



Research paper

Parasite–host network analysis provides insights into the evolution of two mistletoe lineages (Loranthaceae and Santalaceae)

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ABSTRACT

Mistletoes are ecologically important parasitic plants, with > 1600 species from five lineages worldwide. Mistletoe lineages exhibit distinct patterns of species diversification and host specificity, however, the mechanisms underlying these differences are poorly understood. In this study, we analysed a comprehensive parasite–host network, including 280 host species from 60 families and 22 mistletoe species from two lineages (Santalaceae and Loranthaceae) in Xishuangbanna, located in a biodiversity hotspot of tropical Asia. We identified the factors that predict the infection strength of mistletoes. We also detected host specificity and the phylogenetic signal of mistletoes and their hosts. We found that this interaction network could be largely explained by a model based on the relative abundance of species. Host infection was positively correlated with diameter at breast height and tree coverage, but negatively correlated with wood density. Overall, closely related mistletoe species tend to interact more often with similar hosts. However, the two lineages showed a significantly different network pattern. Rates of host generality were higher in Loranthaceae than in Santalaceae, although neither lineage showed phylogenetic signal for host generality. This study demonstrates that the neutral interaction hypothesis provides suitable predictions of the mistletoe–host interaction network, and mistletoe species show significant phylogenetic signals for their hosts. Our findings also indicate that high species diversification in Loranthaceae may be explained by high rates of host generality and the evolutionary history shared by Loranthaceae species with diverse host plants in the tropics.

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1. Introduction

The complexity of natural communities stems from the huge number of species present in a community and the diversity and unequal strengths of the interactions among them (Dunne et al., 2002; Bascompte, 2010). In recent decades, interactions among species along the mutualism to antagonism continuum have been

defined as networks, which provide an ideal framework for investigating community-level patterns of species interactions, as well as community structure, dynamics, and stability (Thompson and Fernandez, 2006; Tylianakis et al., 2010; Delmas et al., 2019). This framework may provide new insights to better understand the evolutionary history of species (Tobias et al., 2014; Allio et al., 2021).

One of the aims of community ecology is to recognize the factors driving the establishment of species interactions. Quantitative network analysis of mutualism, antagonism, and commensalism has revealed general patterns of community organisation. One example is nestedness, which refers to the concept that more specialist species tend to interact with a subset of species from the partners of generalised species (Bascompte et al., 2003). Furthermore, the differences in network characters (e.g., nestedness)

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among symbiotic interactions indicate different mechanisms underlying the interacting structures (Fontaine et al., 2009; Thébaud and Fontaine, 2010; Piazzon et al., 2011). For instance, research on host–pathogen networks indicates that nestedness is driven by intraspecific competition, while modularity depends on local adaptation and competition (Valverde et al., 2020). Thus far, many studies have been conducted to detect the factors that affect the network structure of antagonisms (Enquist et al., 1999; Cagnolo et al., 2011; Poisot et al., 2012), mutualisms (Stang et al., 2009; Vázquez et al., 2009a; Donatti et al., 2011; Olesen et al., 2011), and commensalisms (Sáyago et al., 2013; Atencio et al., 2021). Factors determining the structural patterns of networks may be neutral, biological (such as complementary phenotypes), or both (Jordano et al., 2003; Santamaría and Rodríguez-Gironés, 2007; Stang et al., 2007; Vázquez et al., 2009b). Additionally, network structures may result from multi-hierarchical and nonexclusive interactions between factors (Vázquez et al., 2009b).

Parasites are integral elements of natural communities not only because they are important components of biomass and biodiversity (Frainer et al., 2018) but also because they can regulate richness and/or re-order the relative abundance of species and thus alter the community structure (Wood et al., 2007). As a special group of parasitic plants, mistletoes are aerial hemiparasites of the order Santalales and comprise more than 1600 species with five distinctive lineages (Nickrent et al., 2010; APG IV, 2016; Nickrent, 2020). All mistletoe species are obligate hemiparasites, obtaining resources, including minerals and water, from hosts while retaining their capacity to photosynthesize (Nickrent, 2020). Some mistletoes are regarded as keystone species because they promote biodiversity by providing structural (nesting space) and nutritional resources in their ecosystems (Watson and Herring, 2012; Rodríguez-Cabal et al., 2013; Mellado and Zamora, 2017).

Elucidating the interaction intensity and infection dynamics between mistletoes and hosts is fundamental to understanding the ecological processes of this antagonistic interaction. Two main hypotheses have been proposed to interpret the processes and patterns in networks of species interactions. The “neutral hypothesis” assumes that individuals in a community interact randomly; therefore, patterns of network interaction mainly depend on species abundances (Vázquez, 2005; Vázquez et al., 2007, 2009b). In contrast, the “forbidden links hypothesis” (or “barrier model”) proposes that interactions are dependent on matching morphological traits, i.e., ecological trait mismatches constrain (Santamaría and Rodríguez-Gironés, 2007; Stang et al., 2009; Vázquez et al., 2009b; Maruyama et al., 2014; Sazatornil et al., 2016). Host specificity of parasites has been proposed as an evolutionary adaptation, as even generalist parasitic species may display a certain degree of host preference (Bernays and Graham, 1988). In natural communities, the relative abundance of hosts can be a pivotal element in shaping the host preferences of a parasite species (Norton and Carpenter, 1998; Milner et al., 2020). For example, a study of a generalist parasite in South America, *Tristerix corymbosus* (Loranthaceae), showed that its host specificity differed among localities and was affected by host coverage (Atencio et al., 2021). Furthermore, intrinsic host traits, including chemical defences and bark properties, may explain host specificity (Vidal-Russell and Premoli, 2015).

Phylogenetic constraints have also been identified for preferential host use (Atencio et al., 2021). Three categories of parasites can be identified depending on host preference: oixenous, which infect a single (specific) host; stenoxenous, which only infect related species (i.e., within the same genus or family); and generalist, which infect several unrelated host species (i.e., from different genera or families) (Combes, 2001; Barrett and Heil, 2012). Phylogenetic specificity should be considered when classifying host

preference; however, studies have failed to detect phylogenetic signal in the host–mistletoe network structure (Sáyago et al., 2013). Consequently, whether phylogenetic history constrains the specificity or generality of mistletoe remains uncertain.

Among the five lineages of mistletoe, Loranthaceae is the largest family, comprising 76 genera and over 1000 species with a worldwide distribution (Nickrent et al., 2019). Santalaceae comprises more than 500 species distributed mainly in temperate and tropical regions, with seven genera and 33 species found in China (Xia and Gilber, 2003). Santalaceae parasites evolved *ca.* 53 Mya, whereas Loranthaceae evolved approximately 28 Mya (Vidal-Russell and Nickrent, 2008a). Compared to Santalaceae, the more recently-evolved Loranthaceae may have experienced rapid adaptive radiation and speciation at the onset of global cooling (Oligocene), when tropical regions narrowed and gave way to temperate grasslands and deciduous forests (Vidal-Russell and Nickrent, 2008a). One intriguing question is whether the differences in evolutionary history between these two major lineages affect today's mistletoe–host interactions.

In this study, we investigated the network structure of hosts and mistletoes (Loranthaceae and Santalaceae) in Xishuangbanna, Yunnan Province, China. We identified the factors that predict infection strength of mistletoes and whether the phylogeny of mistletoes and their hosts constrained interaction patterns. In addition, we assessed host specificity and the phylogenetic signal of mistletoes from both Loranthaceae and Santalaceae to explore how the evolutionary history of the families may affect the interactive ecological network of mistletoes and their hosts.

2. Materials and methods

2.1. Study site

The Xishuangbanna Dai Autonomous Prefecture in Yunnan Province, southwest China, is a pivotal region along the Lancang–Mekong River that connects China and the lower Mekong countries. It is in a biodiversity hotspot of tropical Asia (Fig. S1) with mean annual temperature and precipitation of approximately 21 °C and 1500 mm, respectively (Zhu et al., 2006). The climate of this region is affected by warm, wet air masses from the Indian Ocean in summer (including monsoons), and subtropical continental air masses in winter, resulting in a rainy season (May to October) and a dry season (November to April) (Zhang, 1963). The vegetation types in the study area mainly include tropical monsoon forest, tropical montane evergreen broad-leaved forest, tropical seasonal moist forest, montane rainforest, bamboo forest, shrubland, and grassland (He et al., 2008). More than 20 mistletoe species, including both generalists and specialists, occur in Xishuangbanna, with different spatial distributions (Sui and Zhang, 2014). The recorded host species mainly include members of Euphorbiaceae, Leguminosae, and Moraceae along with representatives of several other families (Wang and Zhang, 2017).

2.2. Host and mistletoe databases

Both geographical distribution and phylogenetic history may drive host specificity and can be investigated in a geographical area (Poulin et al., 2011). We conducted an extensive survey of trees infected by mistletoe in Xishuangbanna between 2015 and 2017. Combining pre-existing trails within an area of 1,912,450 ha, we designed a new set of transects and plots that covered mistletoe habitats. Habitats in the survey included gallery forest, rainforest edges, plantations, and roadsides, with elevations ranging from 490 to 1711 m. The forest sample plots were distributed in nature reserves (including the National Nature Reserve of Xishuangbanna

and the National Nature Reserve of the Naban River), and some block forests protected by local Dai villages. Plantations included parks, tree farms, ecological tea gardens, and botanical gardens (Wang and Zhang, 2017). A new set of 51 transects with areas ranging from $10 \times 10 \text{ m}^2$ to $20 \times 3000 \text{ m}^2$ with intervals $> 200 \text{ m}$ were surveyed (Fig. S1; Table S1). We observed infected hosts by binoculars. Hosts were then recorded and identified. Mistletoe species and number of parasite individuals per infected host were also recorded. Furthermore, diameter at breast height (DBH, cm), tree coverage (TC, m^2), tree height (TH, m), and bark roughness (BR) were calculated for all trees (including infected and uninfected potential hosts) in the plots. The DBH was directly measured for each tree using a DBH ruler (accuracy = 1 mm). TH was measured by laser rangefinder and TC was estimated by planimetric area of tree canopy on the ground. BR was divided into four categories: smooth, rough with no obvious cracks, rough with cracks, rough with deep cracks (Mistry, 1998). Wood density (WD, g/cm^3) was obtained from the Global Wood Density Database (Zanne et al., 2009). Voucher specimens of each mistletoe species were stored at the Herbarium of the Xishuangbanna Tropical Botanical Garden.

2.3. Bipartite network analysis

In this study, qualitative and quantitative networks were used to characterise the patterns of interactions between mistletoes and their host species. The qualitative bipartite matrix included the interactions between mistletoes and their hosts, whereas the quantitative matrix was based on the infection strength of mistletoes for each host species. In the qualitative bipartite network, matrix R , $r_{ij} = 1$ when mistletoe j infected host plant i ; otherwise, $r_{ij} = 0$ (Jordano, 1987; Bascompte et al., 2003). Infection strength was quantified as the infection intensity, i.e., the number of individuals of each mistletoe species per host individual. We calculated network metrics related to the antagonistic metrics, including total links, links per species, connectance, modularity, nestedness, generality, and vulnerability using the *bipartite* package (Dormann et al., 2008, 2009). These metrics included.

- Connectance (C): Realised proportion of possible links in the network, $C = L/(HM)$. L indicates sum of links, H indicates number of host species, M indicates number of mistletoe species.
- H_2' : A quantitative measure of specialisation, ranging between 0 (no specialisation) and 1 (complete specialisation) (Blüthgen et al., 2006).
- ISA: Interaction strength asymmetry for mistletoe and hosts, a value of zero indicates highly symmetric interaction strength, whereas a value close to 1 or -1 indicates high asymmetry.
- Host species richness: the number of host species involved in the network.
- Weighted NODF: Weighted nestedness a quantitative index for nestedness building on NODF, ranging 0 (no nestedness) and 100 (perfect nestedness).
- Generality: weighted mean number of host species per mistletoe species.
- Vulnerability: weighted mean number of mistletoe species per host species.

Two parameters – the within-module degree (Z) and the between-module connectivity (C) – were calculated to identify the species' role (i.e., module hubs, network hubs, peripherals, and connectors) in the network. Measurements of Z and C , and the criteria for classifying the species' role in the network were defined as in Guimerà and Amaral (2005), Olesen et al. (2007), and Ulrich et al. (2009).

Modularity was estimated with the index M ranging 0 and 1 based on a simulated annealing algorithm using the program NetCarto (Guimerà and Amaral, 2005). Its significance was calculated with a Monte Carlo procedure based on 1000 randomizations.

Compared with specialist interactions, generalist interactions will encounter more bias from their increased likelihood. Thus, we calculated the host specialization/generalization index (G) according to Medan et al. (2006). This index ranges from 0 (highly specialist) to 1 (highly generalist).

2.4. Ecological models of network determinants

The method and conception proposed by Vázquez et al. (2009b) were utilized in our study, in which the real observed interaction matrix is considered as a multiple interaction probability matrix resulting from different factors. Multiple interaction probability matrices were simulated for each factor, including the relative abundance of species (A), spatial overlap (S , absence/presence of mistletoes and hosts at each site), and functional traits of the interacting species (DBH, TH, TC, BR, and WD), or for different possible combinations from two or more factors following Sáyago et al. (2013).

Host functional traits (DBH, TH, TC, BR, and WD) and infection intensity were \lg - or square root-transformed to fit normal distribution. Because a strong correlation between DBH and TH ($r = 0.77$, $P < 0.001$) was observed, four functional traits (DBH, TC, BR and WD) were included in the following analysis. To evaluate the predictive power of host functional traits on the interaction network, Generalized Linear Mixed-Effects Models (GLMM) were constructed in the R package *lme4* to test the influence of fixed effects (host functional traits) and random effects (species, transects) on infection intensity (models were checked for the distribution of residuals, and appropriate error distributions were used).

The results of GLMM (Table S2) led us to include three host functional traits (DBH, TC, and WD) in the simulated interaction probability matrix analysis. In total, 32 probability matrices were built (e.g., model A was simulated by abundance of species, model AD by joint factors between species abundance and DBH), in which I rows and J columns correspond to the mistletoe and host species in the network, respectively. All the probability matrices are standardized so that all its elements sum to one (Sáyago et al., 2013).

We considered two methods to select the best model. First, a random method involved running 1000 simulated networks with the bipartite package. In this randomization algorithm, the observed interaction number (sum of the links in matrix Y) was randomly assigned to each cell according to the probability defined by different factors and their assembly. Eight additional indices, weighed NODF, connectance, modularity, generality, H_2' , number of hosts, interaction asymmetry, and vulnerability were calculated for the 1000 simulated networks; the 95% confidence interval was determined for comparisons with indices of the real network (Vázquez et al., 2009b). Probability likelihood analysis was used as an additional method to investigate which of the above drivers offered the best fit for predicting the interactions between mistletoes and their hosts, as well as their strength, in the real networks. Following Vázquez et al. (2009b), the delta AIC (ΔAIC) was used to evaluate the power of each probability model to predict the likelihood of mistletoe–host interactions. A data set of the mistletoe and host (including uninfected potential host) network was extracted because of the WD deficiency; overall, 207 plant species and 22 mistletoe species were included in the multiple interaction probability network analysis. Network statistics were calculated including uninfected woody species in transects, while allowing simulations of a scenario in which all woody species can potentially be colonized by mistletoes and testing whether the

factors considered explain the absence of mistletoe on certain woody species.

2.5. Phylogenetic reconstruction

Phylogenetic relationships among mistletoe species were inferred using nuclear and chloroplast DNA regions. Voucher specimens were deposited at the Xishuangbanna Tropical Botanical Garden herbarium. We sampled 22 mistletoe species from Loranthaceae and Santalaceae. The voucher information and GenBank accession numbers are listed in Table S3. *Erythralium scandens* was used as the outgroup species.

Genomic DNA was extracted from the silica gel-dried tissues or herbarium materials using the cetyltrimethylammonium bromide procedure (Doyle and Doyle, 1987). Polymerase chain reactions and sequencing were performed using the primers designed by Taberlet et al. (1991) and Vidal–Russell and Nickrent (2008). Sequences were aligned using GeneiousPrime 2020.2.4 (<https://www.geneious.com>, Biomatters, Ltd., Auckland, New Zealand; Kearse et al., 2012). We conducted further phylogenetic analysis of the combined data set of three DNA regions (ITS, rbcL, trnL-F) using Bayesian inference (BI). The BI tree was generated based on the general time reversible with a proportion of invariable sites and gamma distribution rate (GTR + I + G) substitution model with 1000 bootstrap replicates. The Akaike information criterion (AIC) was used to determine the best-fitting model for the combined data set, as implemented in jModelTest v.2.1.6 (Darrriba et al., 2012). The phylogenetic tree of mistletoe species was used to confirm the accuracy of placement of the 22 mistletoe species identified in this study (Vidal–Russell and Nickrent, 2008b; Nickrent and Malécot, 2008; Nickrent et al., 2010). For species represented in multiple studies, the topology of the tree generated in this study was congruent with previous work (Der and Nickrent, 2008; Vidal–Russell and Nickrent, 2008b; Nickrent et al., 2010).

The *plantlist* package of R (<https://www.r-project.org>) was used to verify the scientific names of species and *V.PhyloMaker2* package was used to reconstruct the phylogenetic relationships of host species based on the updated mega-phylogeny of plants (Jin and Qian, 2022) and Angiosperm Phylogeny Group (APG) IV (<http://www.theplantlist.org/>). A total of 280 species were included in the host phylogenetic tree for subsequent analyses.

2.6. Evolutionary analysis of mistletoe–host network

We estimated the phylogenetic signal for the Z and C values of each mistletoe and host species (uninfected potential hosts were excluded) using Pagel's lambda (λ) value in the *picante* package (Kembel et al., 2010). We also estimated the phylogenetic signal of G using Pagel's λ . This was calculated based on Pagel's phylogenetic dependence estimator using the *geiger* package to test if either the mistletoes, hosts, or both tended to be distributed in the same module (Pagel, 1999), which can either be derived from a Brownian motion model ($\lambda = 1$) or random model ($\lambda = 0$) (Cooper et al., 2010). The significance of the phylogenetic signal was tested using the log-likelihood ratio and was compared with 999 simulated random models (Cooper et al., 2010).

The Mantel test was used to test the relationship between similarities in the observed interaction matrix and phylogenetic distance among host and mistletoe respectively, using the “mantel function” in the *ecodist* package (Goslee and Urban, 2007; Rezende et al., 2007). For the Mantel test, a significance test of the association between the matrices was based on 10,000 randomised permutations.

3. Results

3.1. General information on bipartite networks

A total of 473 unique links comprising 22 mistletoe species (2 families and 10 genera) and 280 host species (60 families and 180 genera, 270 host species for Loranthaceae mistletoe, and 24 host species for Santalaceae) with connectance (0.08) were recorded in the survey. Different types of host plants, including trees, shrubs, lianas, and herbs, of exotic and native species, were identified. The most parasitised plants were trees (89.3%) followed by shrubs (8.6%) and lianas (1.8%). Network level analyses revealed that mistletoes and hosts were significantly nested (weighted NODF = 20.89, $P < 0.001$), as can be seen in Fig. 1. The index of modularity ($M = 0.42$, $P < 0.001$) indicates the network was modular. We also detected significant specialisation ($H_2' = 0.68$, $P < 0.001$) and high asymmetry ($ISA = -0.44$, $P < 0.001$). The generality of mistletoe is 56.33, while vulnerability for host is 1.64.

Mistletoe species at higher trophic levels, such as *Phacellaria caulescens* and *P. rigidula*, also parasitised other mistletoe species, including *Dendrophthoe pentandra* and *Helixanthera parasitica*. *D. pentandra* was recorded as the most generalist, with 207 host species identified, followed by *Scurrula chingii* var. *yunnanensis*, *Macrosolen cochinchinensis*, *S. chingii*, and *H. parasitica*, with 78, 41, 36, and 33 host species, respectively. The mean number of mistletoe individuals per host was 15.1 ± 1.8 .

The mistletoe species contributed differently to the network (Fig. 1). Six mistletoe species were module hubs (*D. pentandra*, *M. cochinchinensis*, *H. parasitica*, *S. chingii*, *S. chingii* var. *yunnanensis*, and *Scurrula parasitica* var. *graciliflora*), whereas the other mistletoe species were peripheral; most mistletoe species from Santalaceae were peripheral. Furthermore, 22 host species were connectors, while others were peripheral, as displayed in Fig. 1.

Furthermore, host generality, as indicated by the G value, was only marginally higher in Loranthaceae than in Santalaceae ($F = 5.79$, $df = 21$), but this difference was statistically significant ($P = 0.026$, Fig. 2).

3.2. Determinants predicting interaction networks

Generalized Linear mixed model results demonstrated that DBH and TC were positively correlated with host infection (Fig. S2) ($R^2 = 0.09$, $P < 0.001$ for DBH; $R^2 = 0.08$, $P < 0.001$ for TC) but negatively correlated for WD ($R^2 = 0.01$, $P < 0.001$). Of the 32 models adapted to predict the observed interacting matrix between mistletoe and host species, the relative species abundance (A) model was the best fit to predict the interaction network according to the ΔAIC value (Fig. 3). When the network structure was compared, abundance model (A) and joint model including abundance and spatial overlap (AS) in general provided better prediction than other models (Figs. 4 and S3).

Different models were identified for predicting network indices (Fig. 5). The models jointly including abundance, DBH, spatial overlap and wood density (ADSW) performed well for the weighed NODF (Fig. 5). For other network indices such as connectance, modularity, generality, vulnerability, ISA, H_2' and number of host species, models joined with abundance (A) supplied better predictions than other models, indicating abundance contributed the most to explaining the observed patterns (Figs. 5 and S4).

3.3. Evolutionary analysis of mistletoe–host network

Strong phylogenetic signal was detected for mistletoe species in the entire mistletoe–host network (Mantel value = 0.32, z score = 4.64, $P = 0.001$; Table 1). Only the C value showed a strong

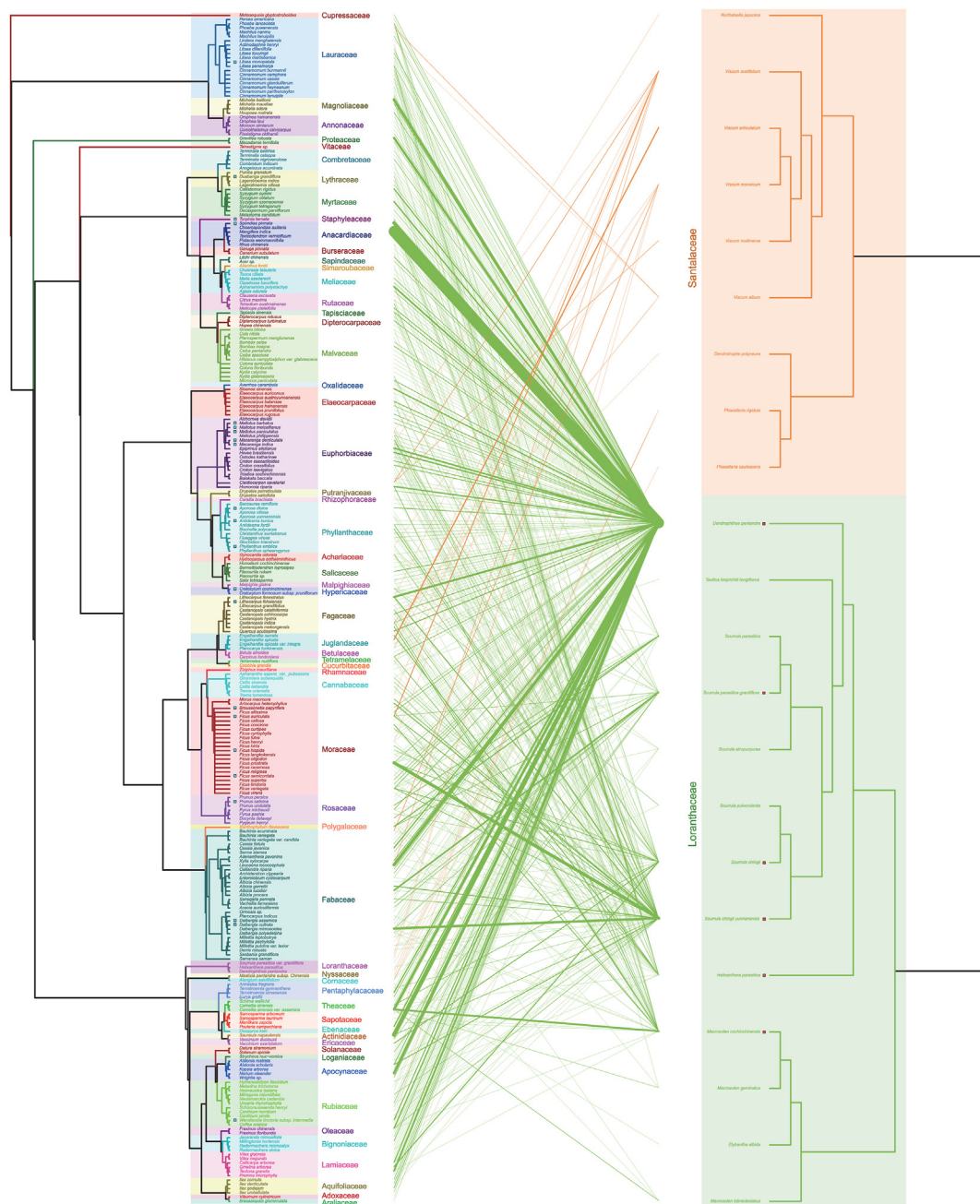


Fig. 1. Bipartite interaction between host and mistletoe species (Note: this is a sub-data set because some of the host species were missing in the published meta-phylogenetic tree). Left, host; right, mistletoe; link width, interaction strength. The red square between branch and species indicates the module hub, while the blue square indicates the connector. Species without squares indicate the peripheral. Colors highlighting species indicate families of host and mistletoe.

phylogenetic signal for mistletoe species in the sub-data from the Loranthaceae networks (Table S4), whereas both the C and Z values of host and mistletoe species in the entire mistletoe–host network and Santalaceae network failed to indicate phylogenetic signals. Phylogenetic signal was not detected for G values of the 22 mistletoe species or the two lineages (Loranthaceae and Santalaceae) (Table S4).

4. Discussion

Our study included a comprehensive overview of mistletoe–host plant interactions, including 280 host species, 22 mistletoe species, and 473 links in a biodiversity hotspot. Generally,

the relative abundance of species and the spatial distributions of mistletoe and host plants provided the best prediction of the interaction network, suggesting that the neutral interaction hypothesis is applicable for predicting this antagonistic network. Furthermore, mistletoe species had significant phylogenetic signals for their hosts. However, there were significant differences between the two main lineages (Loranthaceae and Santalaceae). Specifically, Loranthaceae presented higher host generality than Santalaceae but no phylogenetic signal was observed lined to this generality. This interesting pattern may shed light on how evolutionary history shapes the current interaction network of mistletoes and their hosts.

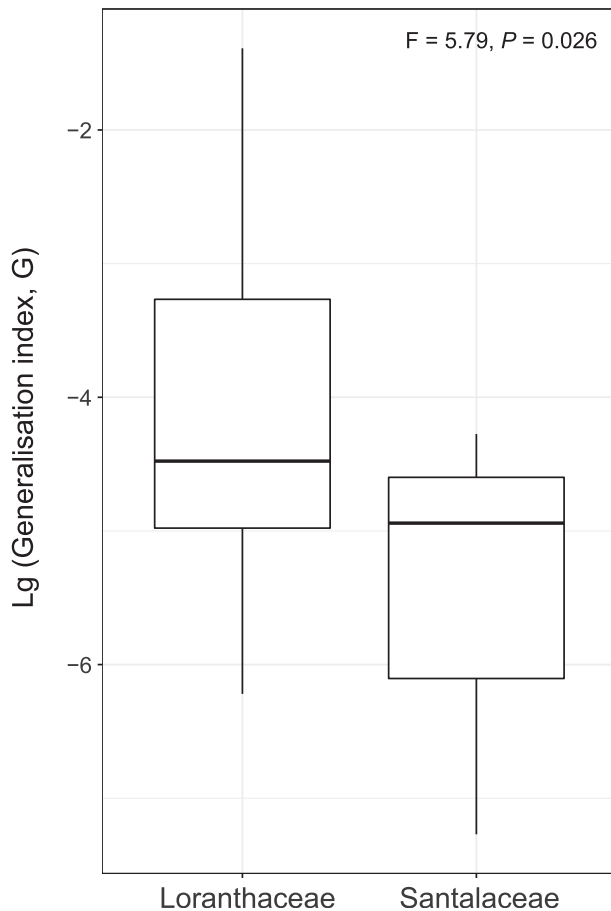


Fig. 2. Host generalisation for Loranthaceae and Santalaceae mistletoe species, as observed in the parasite–host network analysis conducted in Xishuangbanna, Yunnan Province, China, from 2015 to 2017.

4.1. Determinants predicting interaction networks

It is also obvious that biological factors such as the TC, DBH, and WD of hosts affect the establishment of mistletoe–host interactions. For instance, larger trees might provide more spatial and structural capacity for birds (Roxburgh and Nicolson, 2008; Rawsthorne et al., 2011), which in turn allows higher mistletoe densities (Reid and Smith, 2000; Aukema and Martinez del Rio, 2002; Roxburgh and Nicolson, 2008). Other functional traits unassociated with infection such as leaf traits were unable to predict the mistletoe–host network structure (Blick et al., 2012).

In addition, the abundance of species and their spatial distribution also contributed to predicting network structure and the strength of mistletoe–host pairwise interactions in this epiphytic commensalism. Blick et al. (2013) found that mistletoes clump on few trees and interaction strength between mistletoes and hosts are independent on the number of trees. However, they surveyed only a set number of woody trees instead of all trees in transects and only four mistletoe species were detected in the network; therefore, the contribution from abundance of trees might be underestimated. Our results are consistent with previous findings for mistletoes (Norton and Carpenter, 1998; Fadini, 2011; Sáyago et al., 2013), and our results also indicate that neutrality contributes to the network structure to a certain extent (Vázquez et al., 2007; Stang et al., 2009). That is, a network in which assembly includes both abundance and spatial overlap indicates that random encounters of individuals (no matter what species they

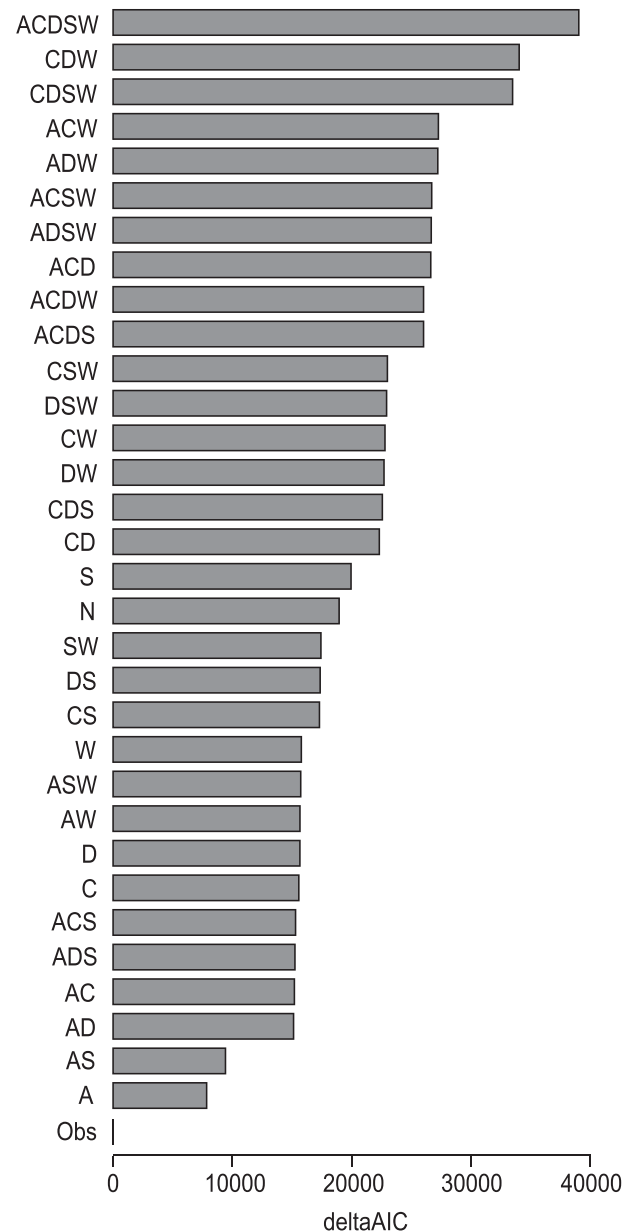


Fig. 3. Delta Akaike Information Criterion (ΔAIC) values for the model determinants of parasite–host network structure, incorporating data on species abundance (model A), tree coverage (model C), diameter at breast height (model D), wood density (model W), spatial overlap (model S), observed data (model Obs), and null model (model N).

belong to) concurring in space contribute largely to the network structure.

In addition, ecological differences among epiphytic species, such as the specific difference in their susceptibility to nutrients, light, or humidity levels, and substrate suitability, may cause differentiation in host preference. Furthermore, mistletoe–host interactions may be influenced by other factors, including seed dispersal, which affects the spatial distribution of mistletoes (Muller-Landau et al., 2002; Pequeno et al., 2022). One study has examined how different frugivorous assemblages of mistletoes alters dominance of mistletoes and, consequently, creates a non-nested and modular pattern of antagonistic interaction between mistletoes and hosts (Genini et al., 2012). However, our study did not investigate the effect of these factors on mistletoe–host interactions in Loranthaceae and Santalaceae at Xishuangbanna.

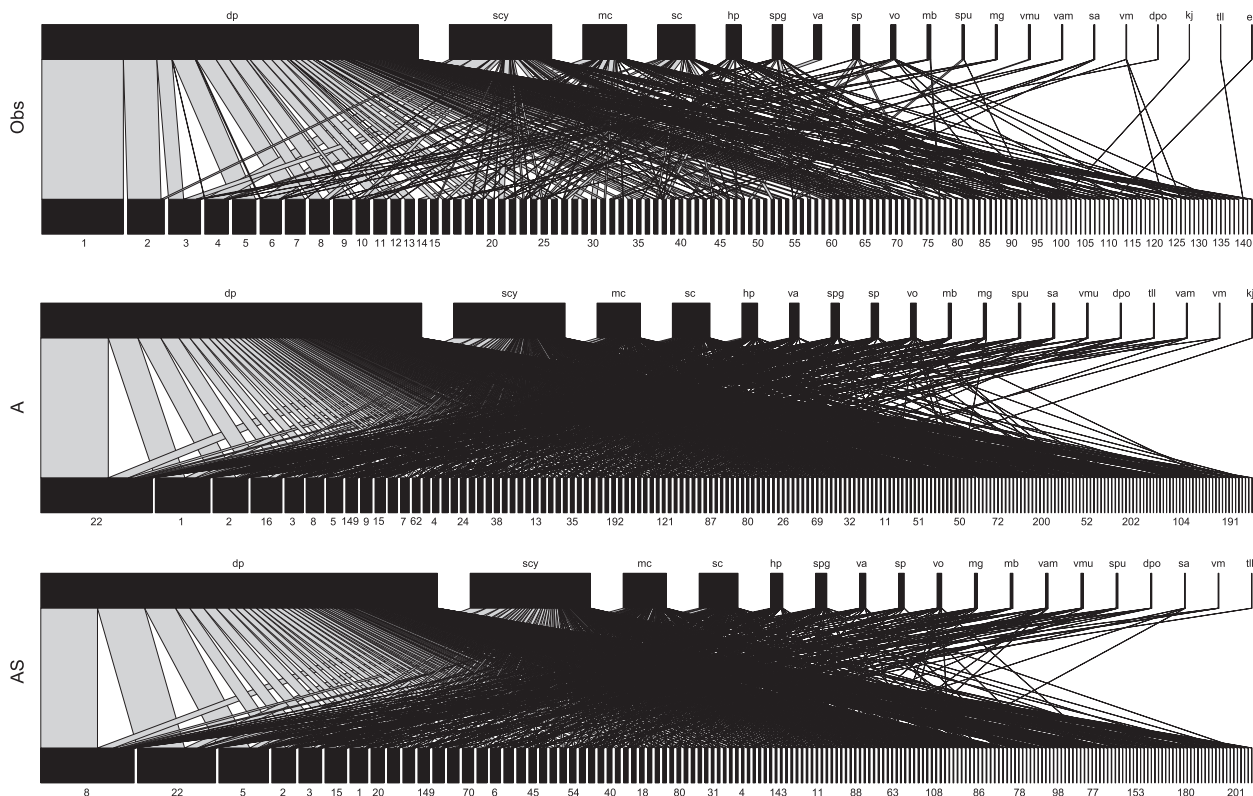


Fig. 4. Observed mistletoe–host network and the network established by one run of the best-performing model of the network determinants abundance (model A) and combined model AS (abundance*spatial overlap). Widths of rectangles represent the relative frequencies of mistletoe–host species interactions. Lines and their width represent mistletoe–host interactions and their relative frequency. Networks are drawn to the same scale. Species codes and order are the same in all networks. Species identity and codes are listed in [Appendix A, Table A5](#).

4.2. Evolutionary analysis of mistletoe–host network

We found that closely related mistletoe species parasitise similar host species. The asymmetry of phylogenetic conservatism of interactions are common and often differs between trophic levels in the same network (Bergamini et al., 2017; Maliet et al., 2020). For example, phylogenetic signals are ubiquitous in plant–pollinator and plant–herbivore networks (Rezende et al., 2007; Edger et al., 2015; Fontaine and Thébaud, 2015; Gilbert et al., 2015; Aizen et al., 2016; Ibanez et al., 2016; Hutchinson et al., 2017; Cirtwill et al., 2020), and previous studies have indicated that these networks often display asymmetric phylogenetic signals. For example, stronger phylogenetic signals have been found for pollinators than for plants, while herbivorous insects showed weaker signals than the plants they feed on (Fontaine and Thébaud, 2015; Aizen et al., 2016; Vitória et al., 2018). Our study indicates that the phylogeny of mistletoe drives the evolution of the mistletoe–host network, implying that infection-related traits are more conservative than the defence traits of the host plant. In antagonistic interactions, the phylogenetic conservatism is frequently stronger at lower trophic (resource) levels than higher trophic (consumer) level (Krasnov et al., 2012; Elias et al., 2013; Fontaine and Thébaud, 2015), suggesting that negative indirect interactions promote a shift in interaction partners and lower conservatism for resource level. For example, the unique mucilaginous viscin that coats seeds contributes to their attachment to woody branches (Muche et al., 2022). Furthermore, closely related hosts co-occur in the same module more often than expected by chance; in contrast, this is rarely the case for parasites (Rezende et al., 2009; Krasnov et al., 2012), and our results show a similar pattern. This pervasive asymmetric pattern implies that the

antagonistic network could be largely shaped by competition, limiting resource overlap among closely related consumer species (Elias et al., 2013). Mistletoe species depend on host species to survive, resulting in strong competition for limited resources, thereby facilitating the plasticity of mistletoe and host interactions.

4.3. Host generality of Loranthaceae vs. Santalaceae

Mistletoes are obligate hemiparasites that depend on hosts for resources, including water and nutrients, and vary in their host infection preference or specificity (Press and Phoenix, 2005; Okubamichael et al., 2016). Most mistletoe species infect only one or a small number of hosts, penetrating their vascular systems to obtain water and nutrients. Consequently, host specificity likely results from evolutionary adaptations to circumvent the defensive systems of particular host species.

In our study, Santalaceae and Loranthaceae mistletoes share a similar region, and both have specialist and generalist representatives. There was also some degree of overlap in host usage observed for members of these two families (Fig. 1). However, we found that the host specificity of Santalaceae was significantly higher than that of Loranthaceae. We propose several hypotheses to explain this pattern. Firstly, previous studies have suggested that mistletoes distributed in diverse regions tend to be host generalists; in contrast, mistletoes in barren regions tend to be specialists (Barlow and Wiens 1977; Norton and Carpenter, 1998). However, Amico et al. (2019) demonstrated that mistletoes with large geographic ranges also have large host ranges, but less diverse regions have more specialized mistletoes. Loranthaceae mistletoes are distributed worldwide (Nickrent et al., 2019), whereas Santalaceae occur mainly in temperate regions, with some representatives in the

Appendix A. Supplementary data

Supplementary data to this article can be found online at <https://doi.org/10.1016/j.pld.2023.03.008>.

References

- Aizen, M.A., Gleiser, G., Sabatino, M., et al., 2016. The phylogenetic structure of plant–pollinator networks increases with habitat size and isolation. *Ecol. Lett.* 19, 29–36.
- Amico, G.C., Nickrent, D.L., Vidal-Russell, R., 2019. Macroscale analysis of mistletoe host ranges in the Andean–Patagonian forest. *Plant Biol.* 21, 150–156.
- Allio, R., Nabholz, B., Wanke, S., et al., 2021. Genome-wide macroevolutionary signatures of key innovations in butterflies colonizing new host plants. *Nat. Commun.* 12, 354.
- APG IV, 2016. An update of the Angiosperm Phylogeny Group classification for the orders and families of flowering plants: APG IV. *Bot. J. Linn. Soc.* 181, 1–20.
- Atencio, N.O., Vidal-Russell, R., Chacoff, N., et al., 2021. Host range dynamics at different scales: host use by a hemiparasite across its geographical distribution. *Plant Biol.* 23, 612–620.
- Aukema, J.E., Martinez del Rio, C., 2002. Variation in mistletoe seed deposition: effects of intra- and interspecific host characteristics. *Ecography* 25, 139–144.
- Barlow, B.A., Wiens, D., 1977. Host–parasite resemblance in Australian mistletoes: the case for cryptic mimicry. *Evolution* 31, 69–84.
- Barrett, L.G., Heil, M., 2012. Unifying concepts and mechanisms in the specificity of plant–enemy interactions. *Trends Plant Sci.* 17, 282–292.
- Bascompte, J., Jordano, P., Melián, C.J., et al., 2003. The nested assembly of plant–animal mutualistic networks. *Proc. Natl. Acad. Sci. U.S.A.* 100, 9383–9387.
- Bascompte, J., 2010. Structure and dynamics of ecological networks. *Science* 329, 765–766.
- Bergamini, L.L., Lewinsohn, T.M., Jorge, L.R., 2017. Manifold influences of phylogenetic structure on a plant–herbivore network. *Oikos* 126, 703–712.
- Bernays, E., Graham, M., 1988. On the evolution of host specificity in phytophagous arthropods. *Ecology* 69, 886–892.
- Blick, R.A.J., Burns, K.C., Moles, A.T., 2012. Predicting network topology of mistletoe–host interactions: do mistletoes really mimic their hosts? *Oikos* 121, 761–771.
- Blick, R.A.J., Burns, K.C., Moles, A.T., 2013. Dominant network interactions are not correlated with resource availability: a case study using mistletoe host interactions. *Oikos* 122, 889–895.
- Blüthgen, N., Menzel, F., Blüthgen, N., 2006. Measuring specialization in species interaction networks. *BMC Ecol.* 6, 9.
- Cagnolo, L., Salvo, A., Valladares, G., 2011. Network topology: patterns and mechanisms in plant–herbivore and host–parasitoid food webs. *J. Anim. Ecol.* 80, 342–351.
- Cirtwill, A.R., Riva, G.V.D., Baker, N.J., et al., 2020. Related plants tend to share pollinators and herbivores, but strength of phylogenetic signal varies among plant families. *New Phytol.* 226, 909–920.
- Combes, C., 2001. *Parasitism: the Ecology and Evolution of Intimate Interactions*. Chicago Press, Chicago USA, p. 53.
- Cooper, N., Jetz, W., Freckleton, R.P., 2010. Phylogenetic comparative approaches for studying niche conservatism. *J. Evol. Biol.* 23, 2529–2539.
- Darriba, D., Taboada, G.L., Doallo, R., et al., 2012. jModelTest 2: more models, new heuristics and parallel computing. *Nat. Methods* 9, 772–772.
- Delmas, E., Besson, M., Brice, M.H., et al., 2019. Analysing ecological networks of species interactions. *Biol. Rev.* 94, 16–36.
- Der, J.P., Nickrent, D.L., 2008. A molecular phylogeny of Santalaceae (Santalales). *Syst. Bot.* 33, 107–116.
- Donatti, C.I., Guimarães, P.R., Galetti, M., et al., 2011. Analysis of a hyperdiverse seed dispersal network: modularity and underlying mechanisms. *Ecol. Lett.* 14, 773–781.
- Dormann, C.F., Gruber, B., Fründ, J., 2008. Introducing the bipartite package: analysing ecological networks. *R News* 8, 8–11.
- Dormann, C., Fründ, J., Blüthgen, N., et al., 2009. Indices, graphs and null models: analyzing bipartite ecological networks. *Open Ecol. J.* 2, 7–24.
- Doyle, J.J., Doyle, J.L., 1987. A rapid DNA isolation procedure for small quantities of fresh leaf tissue. *Phytochem. Bull.* 19, 11–15.
- Dunne, J.A., Williams, R.J., Martinez, N.D., 2002. Network structure and biodiversity loss in food webs: robustness increases with connectance. *Ecol. Lett.* 5, 558–567.
- Edger, P.P., Heidel-Fischer, H.M., Bekaert, M., et al., 2015. The butterfly plant arms-race escalated by gene and genome duplications. *Proc. Natl. Acad. Sci. U.S.A.* 112, 8362–8366.
- Elias, M., Fontaine, C., Frank, van, Veen, F.J., 2013. Evolutionary history and ecological processes shape a local multilevel antagonistic network. *Curr. Biol.* 23, 1355–1359.
- Enquist, B.J., West, G.B., Charnov, E.L., et al., 1999. Allometric scaling of production and life-history variation in vascular plants. *Nature* 401, 907–911.
- Fadini, R., 2011. Non-overlap of hosts used by three congeneric and sympatric loranthaceous mistletoe species in an Amazonian savanna: host generalization to extreme specialization. *Acta Bot. Bras.* 25, 337–345.
- Fontaine, C., Thébault, E., Dajoz, I., 2009. Are insect pollinators more generalist than insect herbivores? *Proc. R. Soc. B-Biol. Sci.* 276, 3027–3033.
- Fontaine, C., Thébault, E., 2015. Comparing the conservatism of ecological interactions in plant–pollinator and plant–herbivore networks. *Popul. Ecol.* 57, 29–36.
- Frainer, A., McKie, B.G., Amundsen, P.A., et al., 2018. Parasitism and the biodiversity–functioning relationship. *Trends Ecol. Evol.* 33, 260–268.
- Genini, J., Cortes, M.C., Guimarães, P.R., et al., 2012. Mistletoes play different roles in a modular host–parasite network. *Biotropica* 44, 171–178.
- Gilbert, G.S., Briggs, H.M., Magarey, R., 2015. The impact of plant enemies shows a phylogenetic signal. *PLoS One* 10, e0123758.
- Goslee, S.C., Urban, D.L., 2007. The ecodist package for dissimilarity-based analysis of ecological data. *J. Stat. Software* 22, 1–19.
- Guimerà, R., Amaral, L.N., 2005. Functional cartography of complex metabolic networks. *Nature* 433, 895–900.
- He, C.G., Feng, Y., Yang, Y.P., 2008. Research on evolution process and driving factors of forest landscape in Xishuangbanna. *Yunnan Geogr. Environ. Res.* 20, 12–17 (in Chinese, with English abstract).
- Hutchinson, M.C., Cagua, E.F., Stouffer, D.B., 2017. Cophylogenetic signal is detectable in pollination interactions across ecological scales. *Ecology* 98, 2640–2652.
- Ibanez, S., Arène, F., Lavergne, S., 2016. How phylogeny shapes the taxonomic and functional structure of plant–insect networks. *Oecologia* 180, 989–1000.
- Jin, Y., Qian, H., 2022. VPhyloMaker2: an updated and enlarged R package that can generate very large phylogenies for vascular plant. *Plant Divers.* 44, 335–339.
- Jordano, P., 1987. Patterns of mutualistic interactions in pollination and seed dispersal—connectance, dependence asymmetries, and coevolution. *Am. Nat.* 129, 657–677.
- Jordano, P., Bascompte, J., Olesen, J.M., 2003. Invariant properties in coevolutionary networks of plant–animal interactions. *Ecol. Lett.* 6, 69–81.
- Kearse, M., Moir, R., Wilson, A., et al., 2012. Geneious basic: an integrated and extendable desktop software platform for the organization and analysis of sequence data. *Bioinformatics* 28, 1647–1649.
- Kembel, S.W., Cowan, P.D., Helmus, M.R., et al., 2010. Picante: R tools for integrating phylogenies and ecology. *Bioinformatics* 26, 1463–1464.
- Krasnov, B.R., Fortuna, M.A., Mouillot, D., et al., 2012. Phylogenetic signal in module composition and species connectivity in compartmentalized host–parasite networks. *Am. Nat.* 179, 501–511.
- Maliet, O., Loeuille, N., Morlon, H., 2020. An individual-based model for the eco-evolutionary emergence of bipartite interaction networks. *Ecol. Lett.* 23, 1623–1634.
- Maruyama, P.K., Vizentin-Bugoni, J., Oliveira, G.M., et al., 2014. Morphological and spatio-temporal mismatches shape a Neotropical savanna plant–hummingbird network. *Biotropica* 46, 740–747.
- Medan, D., Basilio, A., Devoto, M., et al., 2006. Measuring generalization and connectance in temperate, long-lasting systems. In: Waser, N.M., Ollerton, J. (Eds.), *Plant Pollinator Interactions. From Specialization to Generalization*. University of Chicago Press, Chicago, USA, pp. 245–259.
- Mellado, A., Zamora, R., 2017. Parasites structuring ecological communities: the mistletoe footprint in Mediterranean pine forests. *Funct. Ecol.* 31, 2167–2176.
- Milner, K.V., Leigh, A., Gladstone, W., et al., 2020. Subdividing the spectrum—quantifying host specialization in mistletoes. *Botany* 98, 533–543.
- Mistry, J., 1998. Corticolous lichens as potential bioindicators of fire history: a study in the Cerrado of the Distrito Federal, central Brazil. *J. Biogeogr.* 25, 409–441.
- Muller-Landau, H.C., Wright, S.J., Calderon, O., et al., 2002. Assessing recruitment limitation: concepts, methods and case-studies from a tropical forest. In: Levey, D.J., Silva, W.R., Galetti, M. (Eds.), *Seed Dispersal and Frugivory: Ecology, Evolution and Conservation*. CAB International, pp. 35–53.
- Muche, M., Muasya, A.M., Tsegay, B.A., 2022. Biology and resource acquisition of mistletoes, and the defense responses of host plants. *Ecol. Process.* 11, 24.
- Nickrent, D.L., Malécot, V., 2008. Molecular phylogenetic relationships of Olacaceae and related Santalales. *Syst. Bot.* 33, 97–106.
- Nickrent, D.L., Malécot, V., Vidal-Russell, R., et al., 2010. A revised classification of Santalales. *Taxon* 59, 538–558.
- Nickrent, D.L., Anderson, F., Kuijt, J., 2019. Inflorescence evolution in Santalales: integrating morphological characters and molecular phylogenetics. *Am. J. Bot.* 106, 402–414.
- Nickrent, D.L., 2020. Parasitic angiosperms: how often and how many? *Taxon* 69, 5–27.
- Norton, D.A., Carpenter, M.A., 1998. Mistletoes as parasites: host specificity and speciation. *Trends Ecol. Evol.* 13, 101–105.
- Norton, D.A., De Lange, P.J., 1999. Host specificity in parasitic mistletoes (Loranthaceae) in New Zealand. *Funct. Ecol.* 13, 552–559.
- Olesen, J.M., Bascompte, J., Dupont, Y.L., et al., 2011. Missing and forbidden links in mutualistic networks. *Proc. R. Soc. B-Biol. Sci.* 278, 725–732.
- Olesen, J.M., Bascompte, J., Dupont, Y.L., et al., 2007. The modularity of pollination networks. *Proc. Natl. Acad. Sci. U.S.A.* 104, 19891–19896.
- Okubamichael, D.Y., Griffiths, M.E., Ward, D., 2016. Host specificity in parasitic plants—perspectives from mistletoes. *AoB Plants* 8, plw069.
- Pagel, M., 1999. Inferring the historical patterns of biological evolution. *Nature* 401, 877–884.
- Pequeno, P.A.C.L., Franklin, E., Norton, R.A., 2022. Modelling selection, drift, dispersal and their interactions in the community assembly of Amazonian soil mites. *Oecologia* 196, 805–814.
- Piazzon, M., Larrinaga, A.R., Santamaría, L., 2011. Are nested networks more robust to disturbance? A test using epiphyte–tree, commensalistic networks. *PLoS One* 6, e19637.
- Poisot, T., Thrall, P.H., Hochberg, M.E., 2012. Trophic network structure emerges through antagonistic coevolution in temporally varying environments. *Proc. R. Soc. B-Biol. Sci.* 279, 299–308.

- Poulin, R., Krasnov, B.R., Mouillot, D., 2011. Host specificity in phylogenetic and geographic space. *Trends Parasitol.* 27, 355–361.
- Press, M.C., Phoenix, G.K., 2005. Impacts of parasitic plants on natural communities. *New Phytol.* 166, 737–751.
- Rawsthorne, J., Watson, D.M., Roshier, D.A., 2011. Implications of movement patterns of a dietary generalist for mistletoe seed dispersal. *Austral Ecol.* 36, 650–655.
- Reid, N., Smith, S.S., 2000. Population dynamics of an arid zone mistletoe (*Amyema preissii*, Loranthaceae) and its host *Acacia victoriae* (Mimosaceae). *Aust. J. Bot.* 48, 45–58.
- Rezende, E.L., Lavabre, J.E., Guimarães, P.R., et al., 2007. Non-random coextinctions in phylogenetically structured mutualistic networks. *Nature* 448, 925–928.
- Rezende, E.L., Albert, E.M., Fortuna, M.A., et al., 2009. Compartments in a marine food web associated with phylogeny, body mass, and habitat structure. *Ecol. Lett.* 12, 779–788.
- Rodríguez-Cabal, M.A., Barrios-García, M.N., Amico, G.C., et al., 2013. Node-by-node disassembly of a mutualistic interaction web driven by species introductions. *Proc. Natl. Acad. Sci. U.S.A.* 110, 16503–16507.
- Roxburgh, L., Nicolson, S.W., 2008. Differential dispersal and survival of an African mistletoe: does host size matter? *Plant Ecol.* 195, 21–31.
- Santamaría, L., Rodríguez-Gironés, M.A., 2007. Linkage rules for plant–pollinator networks: trait complementarity or exploitation barriers? *PLoS Biol.* 5, e31.
- Sáyago, R., Lopezaiza-Mikel, M., Quesada, M., et al., 2013. Evaluating factors that predict the structure of a commensalistic epiphyte–phorophyte network. *Proc. R. Soc. B-Biol. Sci.* 280, 20122821.
- Sazatornil, F.D., Moré, M., Benítez-Vieyra, S., et al., 2016. Beyond neutral and forbidden links: morphological matches and the assembly of mutualistic hawkmoth–plant networks. *J. Anim. Ecol.* 85, 1586–1594.
- Stang, M., Klinkhamer, P.G.L., van der Meijden, E., 2007. Asymmetric specialization and extinction risk in plant–flower visitor webs: a matter of morphology or abundance. *Oecologia* 151, 442–453.
- Stang, M., Klinkhamer, P.G.L., Waser, N.M., et al., 2009. Size-specific interaction patterns and size matching in a plant–pollinator interaction web. *Ann. Bot.* 103, 1459–1469.
- Sui, Y., Zhang, L., 2014. Spatial distribution pattern of mistletoe species in tropical plantation. *J. Yunnan Univ.* 36, 755–764 (in Chinese, with English abstract).
- Taberlet, P., Gielly, L., Pautou, G., et al., 1991. Universal primers for amplification of three non-coding regions of chloroplast DNA. *Plant Mol. Biol.* 17, 1105–1109.
- Thébault, E., Fontaine, C., 2010. Stability of ecological communities and the architecture of mutualistic and trophic networks. *Science* 329, 853–856.
- Thompson, J.N., Fernandez, C.C., 2006. Temporal dynamics of antagonism and mutualism in a geographically variable plant–insect interaction. *Ecology* 87, 103–112.
- Tobias, J.A., Cornwallis, C.K., Derryberry, E.P., et al., 2014. Species coexistence and the dynamics of phenotypic evolution in adaptive radiation. *Nature* 506, 359–363.
- Tylianakis, J.M., Laliberté, E., Nielsen, A., et al., 2010. Conservation of species interaction networks. *Biol. Conserv.* 143, 2270–2279.
- Ulrich, W., Almeida-Neto, M., Gotelli, N.J., 2009. A consumer's guide to nestedness analysis. *Oikos* 118, 3–17.
- Valverde, S., Vidiella, B., Montañez, R., et al., 2020. Coexistence of nestedness and modularity in host–pathogen infection networks. *Nat. Ecol. Evol.* 4, 568.
- Vázquez, D.P., 2005. Degree distribution in plant–animal mutualistic networks: forbidden links or random interactions? *Oikos* 108, 421–426.
- Vázquez, D.P., Melián, C.J., Williams, N.M., et al., 2007. Species abundance and asymmetric interaction strength in ecological networks. *Oikos* 116, 1120–1127.
- Vázquez, D.P., Blüthgen, N., Cagnolo, L., et al., 2009a. Uniting pattern and process in plant–animal mutualistic networks: a review. *Ann. Bot.* 103, 1445–1457.
- Vázquez, D.P., Chacoff, N.P., Cagnolo, L., 2009b. Evaluating multiple determinants of the structure of plant? Animal mutualistic networks. *Ecology* 90, 2039–2046.
- Vidal-Russell, R., Nickrent, D.L., 2008a. The first mistletoes: origins of aerial parasitism in Santalales. *Mol. Phylogenet. Evol.* 47, 523–537.
- Vidal-Russell, R., Nickrent, D.L., 2008b. Evolutionary relationships in the showy mistletoe family (Loranthaceae). *Am. J. Bot.* 95, 1015–1029.
- Vidal-Russell, R., Premoli, A.C., 2015. Nothofagus trees show genotype difference that influence infection by mistletoes, Misodendraceae. *Aust. J. Bot.* 63, 541–548.
- Vitória, R.S., Vizenin-Bugoni, J., Duarte, L.D.S., 2018. Evolutionary history as a driver of ecological networks: a case study of plant–hummingbird interactions. *Oikos* 127, 561–569.
- Wang, X., Zhang, L., 2017. Species diversity and distribution of mistletoes and hosts in four different habitats in Xishuangbanna, Southwest China. *J. Yunnan Univ.* 39, 701–711 (in Chinese, with English abstract).
- Watson, D.M., Herring, M., 2012. Mistletoe as a keystone resource: an experimental test. *Proc. R. Soc. B-Bio. Sci.* 279, 3853–3860.
- Wood, C.L., Byers, J.E., Cottingham, K.L., et al., 2007. Parasites alter community structure. *Proc. Natl. Acad. Sci. U.S.A.* 104, 9335–9339.
- Xia, N., Gilber, M.G., 2003. Santalaceae. In: *Flora of China*, vol. 5. Science Press and Missouri Botanical Garden Press, Beijing and St. Louis, pp. 218–219.
- Zanne, A.E., Lopez-Gonzalez, G., Coomes, D.A., et al., 2009. Global wood density database. *Plant Economics. Dryad Digital Reposit.* <https://doi.org/10.5061/dryad.234>.
- Zhang, K., 1963. A preliminary study on the climatic characteristic and the formation factors in southern Yunnan. *Acta Meteorol. Sin.* 33, 218–230 (In Chinese, with English abstract).
- Zhu, H., Cao, M., Hu, H., 2006. Geological history, flora, and vegetation of Xishuangbanna, southern Yunnan, China. *Biotropica* 38, 310–317.